

Original Research

**N<sub>2</sub>O Emission Pattern in A Legume-Based Agroecosystem**

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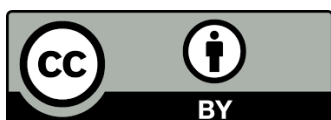
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**Received:** January 26, 2023**Accepted:** April 22, 2023**Published:** April 26, 2023**Abstract**

Legumes provide several ecological services to agroecosystems, but there is a lack of references on services related to N flows for a wide range of legume crops. N<sub>2</sub>O emissions were measured in two field experiments using a two-year legume-cereal crop sequence. In the first year (2014 and 2016), different legume crops were grown (lupin, pea, fava bean, common bean, soybean, chickpea) and compared to fertilized cereals (barley and sorghum). Once the seeds were harvested and the residues incorporated in the soil, unfertilized wheat was sown and harvested in the second year (2015 and 2017). N<sub>2</sub>O emissions, as well as soil temperature and moisture, were measured continuously using an automated chamber method during the two years of each experiment. Daily N<sub>2</sub>O emissions were less than 10 g N-N<sub>2</sub>O ha<sup>-1</sup> d<sup>-1</sup>, with higher values (ranging from 10 to 90 g N-N<sub>2</sub>O ha<sup>-1</sup> d<sup>-1</sup>) being measured during exceptionally rainy conditions. Daily N<sub>2</sub>O emissions were mainly influenced by climatic conditions for field experiments and far less by inorganic N content, except for N-fertilized cereals. For both field experiments, cumulative N<sub>2</sub>O emissions during legume and cereal pre-crops + fallow period between pre-crop harvest and wheat sowing (1<sup>st</sup> year) (mean values 365.4 and 318.1 g N-N<sub>2</sub>O ha<sup>-1</sup> for experiment I and II, respectively) were higher than during wheat crop cultivation (2<sup>nd</sup> year) (155.8 and 101.5 g N-N<sub>2</sub>O ha<sup>-1</sup> for experiment I and II,



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respectively). For field experiment II, N<sub>2</sub>O emissions were slightly higher for the N fertilized cereal pre-crops (529.8 and 523.3 g N-N<sub>2</sub>O ha<sup>-1</sup> for barley and sorghum, respectively) compared to legume pre-crops (mean values 380.6 and 417.2 g N-N<sub>2</sub>O ha<sup>-1</sup> for legumes sown in March and May, respectively), while no significant difference was measured for field experiment I. There was no difference in N<sub>2</sub>O emissions during the cultivation of the different legume species. Furthermore, when wheat was grown after legumes or N fertilized cereals, N<sub>2</sub>O emissions were comparable for the different experimental treatments with no relation established with the amounts of N present in crop residues or their C: N ratios. Despite the small differences in emissions between N-fertilized cereals and grain legumes, introducing these leguminous species in crop rotation and in these pedoclimatic conditions makes it possible to substitute synthetic N fertilizer and mitigate the greenhouse gases emitted from these cropping systems. However, further research is still needed to clarify and quantify the value of legumes in mitigating and reducing greenhouse gas emissions from cropping systems.

### Keywords

Agricultural soil; nitrous oxide; N mineralization; leguminous species; cereal; crop rotation; crop residues

## 1. Introduction

Global warming is associated with increased greenhouse gas (GHG) atmospheric concentrations, predominantly caused by anthropic activities. The AFOLU (Agriculture, Forestry and Other Land Uses) sector is estimated to produce 21% of total net anthropogenic GHG emissions, including 69% of the potent GHG N<sub>2</sub>O emissions [1]. N<sub>2</sub>O is emitted from the soil during the N cycle through two microbial processes: nitrification and denitrification. N<sub>2</sub>O emissions are related to the inorganic N content of agricultural soils [2] and are greatly impacted by inorganic and organic N fertilization.

Legumes are an alternative source of N for cropping systems. Due to their ability to fix atmospheric N via symbiosis with soil bacteria, legumes provide several ecological services to agroecosystems. Their cultivation results in the harvest of proteins or protein-rich fodder, without any N fertilization. In addition, their N-rich crop residues left in the soil can supply N to the following crop and reduce the amount of N fertilizer required. Lastly, their cultivation reduces fossil fuel energy and the GHG emissions associated with using nitrogen fertilizers, as reported by several authors [3, 4].

Previous studies have shown that N<sub>2</sub>O emissions from legume crops without N fertilization are lower than those from non-legume fertilized crops [5-7]. Indeed, the symbiotic fixation of atmospheric nitrogen is recognized as a non-N<sub>2</sub>O emitting process [3, 8]. On the other hand, the degradation of nodules at the end of the crop and the mineralization of the organic nitrogen contained in the crop residues can lead to N<sub>2</sub>O emissions [7, 9]. Thus, contrasting values of N<sub>2</sub>O emissions in cropping systems with or without legumes have been reported in the literature, with emissions being influenced by agricultural practices, particularly tillage, crop rotation and legume cultivation [10-12]. Peaks of N<sub>2</sub>O emissions are generally reported after the incorporation of legume crop residues (e.g., [4, 13, 14]). The lower C/N ratio of legume residues is responsible for their rapid

mineralization in the soil and mineral nitrogen availability for nitrification and denitrification processes, depending on aerobic/anaerobic soil conditions. However, very few of these studies have compared the N<sub>2</sub>O emissions from soil successively cultivated with (1) different species of grain legumes (e.g. [8, 12, 15]) and (2) different species of cereals in the same experimental conditions.

The current study aimed to measure N<sub>2</sub>O emissions during a two-year legume-cereal crop sequence: i) the first year with different legume crops compared to N fertilized cereals (named pre-crops in this study), and ii) the second year with unfertilized winter wheat that had taken up mineralized nitrogen from legume or cereal residues. The first main hypothesis was that N<sub>2</sub>O emissions are lower for legume crops than N-fertilized cereals. The second main hypothesis was N<sub>2</sub>O emissions from a cereal crop are higher when cultivated after incorporating legume residues into the soil than cereal residues. The originality of this study is that it investigated N<sub>2</sub>O emissions for a wide range of legumes during their cultivation and the following crop (i.e., wheat).

## 2. Materials and Methods

### 2.1 Experimental Site

Two field experiments, each lasting two years, were carried out at the INRAE Dijon experimental site located in F 21110 Breteniere (Eastern France; N 47° 14' 27.6'', E 5° 6' 54'', altitude = 206 m a.s.l.) in 2014-2015 (Experiment I) and 2016-2017 (Experiment II). The soil is classified as Eutric Cambisol (Anoclayic) [16] with a clay surface layer (depth = 0.65 ± 0.15 m) developed on a coarse alluvial layer (Table S1).

The site is subject to a semi-continental climate but the climatic conditions were contrasted between Experiments I and II. Cumulative rainfall and mean temperature between the beginning of the pre-crop harvest and wheat harvest were 850 mm and 11.8°C in Experiment I and 600 mm and 11.3°C in Experiment II (Figure S1). Rainfall was unusually high between July and November 2014 (541 mm) compared to 2016 (303 mm) and to the 20-year mean rainfall during the same period (362 mm). Rainfall between January and June was similar in 2015 (252 mm) and 2017 (293 mm), but lower than the 20-year mean rainfall during the same period (340 mm).

### 2.2 Experimental Design

Each experiment consisted of a two-year crop sequence (i.e., legume or cereal pre-crop followed by wheat) as described previously [17]. Legume and cereal pre-crop treatments were applied in 2014, followed by winter wheat harvested in 2015 (Experiment I). In 2016, the same pre-crops were grown near the first experiment, followed by winter wheat harvested in 2017 (Experiment II).

#### 2.2.1 Pre-Crop Treatments

During the first year of both experiments I and II, 8 treatments were performed in the field including six legume pre-crops: fava bean (*Vicia faba*), chickpea (*Cicer arietinum*), common bean (*Phaseolus vulgaris*), lupin (*Lupinus albus*), pea (*Pisium sativum*), and soybean (*Glycine max*) and two N fertilized cereal pre-crops: barley (*Hordeum vulgare*) and sorghum (*Sorghum bicolor*). Pre-crops were sown either in March or in May, depending on the physiology of each species (Table S2), and were called spring pre-crops and summer pre-crops, respectively, after that. Pre-crops were sown on plots 1.5 m wide and 12 m long using four randomly distributed replicates in Experiments I and

II. No N fertilization was applied to any of the legume pre-crops in either experiment. In contrast, N fertilization was applied in the form of  $\text{NH}_4\text{NO}_3$  at rates of  $135 \text{ kg N ha}^{-1}$  in Experiment I (60 and  $75 \text{ kg N}$  on 3<sup>rd</sup> and 28<sup>th</sup> April 2014) and  $70 \text{ kg N ha}^{-1}$  in Experiment II (19<sup>th</sup> April 2016) for barley. N fertilizer was also applied twice during sorghum growth at rates of  $50 \text{ kg N ha}^{-1}$  in 2014 (9<sup>th</sup> May and 13<sup>th</sup> June) and 2016 (7<sup>th</sup> June and 7<sup>th</sup> July). In Experiment I, spring pre-crops was irrigated with 35, 40 and 50 mm on March 26<sup>th</sup>, May 16<sup>th</sup> and June 18<sup>th</sup> 2014, respectively. Summer pre-crops were irrigated with 40, 40 and 50 mm on May 16<sup>th</sup>, May 27<sup>th</sup> and June 14<sup>th</sup> 2014, respectively. In Experiment II, 50 mm was supplied to summer pre-crops on July 11<sup>th</sup> 2016. Weeds were controlled by hand weeding and one insecticide (Karate Zeon) was applied on April 25<sup>th</sup> 2014 and April 21<sup>th</sup> 2016 in Experiments I and II, respectively.

### 2.2.2 Pre-Crop Harvest and Residue Management

Pre-crop grains were harvested with a harvesting machine. No grains were harvested for chickpeas in 2014, due to rainy conditions coupled with excessively low temperatures unsuitable for grain production. In this case, the whole plant was considered crop residue. Two to seven weeks after the pre-crop grain harvest, crop residues were chopped and incorporated into the soil using a rototiller (see Table S2 for pre-crop incorporation dates).

### 2.2.3 Following Crop

Winter wheat (*Triticum aestivum*, cv. Rubisko) was sown on the same plots on October 20<sup>th</sup> 2014 ( $300 \text{ seeds m}^{-2}$ ) in Experiment I and October 24<sup>th</sup> 2016 ( $350 \text{ seeds m}^{-2}$ ) in Experiment II. To control weeds and pathogens, a herbicide (February 26<sup>th</sup> 2015, Arbalète) and a fungicide (April 30<sup>th</sup> 2015, Adexar + Voxan) were applied on wheat in Experiment I. In Experiment II, irrigation was applied on wheat on April 5<sup>th</sup> and 19<sup>th</sup> 2017 (27 mm and 30 mm, respectively) (Figure S1B). In order to better characterize the effect of previous crops on  $\text{N}_2\text{O}$  emissions during wheat cultivation, no N fertilization was applied on wheat.

## 2.3 $\text{N}_2\text{O}$ Emissions Measurements

Nitrous oxide ( $\text{N}_2\text{O}$ ) emissions were measured continuously using automated chambers (length 70 cm, width 70 cm, height 30 cm) with the method described previously [18]. For each treatment the chambers were randomly placed on two of the four plots (i.e., two chambers on one plot and one chamber on the other plot), because of the limited physical connection between the analyzer and chambers. For an experiment I, the chambers were installed on April 16<sup>th</sup> 2014 and 3<sup>rd</sup> June 2014 for pre-crops sown in March and May, respectively. For experiment II, the chambers were installed on 5<sup>th</sup> April 2016 and 6<sup>th</sup> June 2016 for pre-crops sown in March and May, respectively. During the measurement period the chambers were removed for very short periods when harvesting, residue incorporation and wheat sowing were performed. Vegetation was present inside the chambers for the early growth stages of the different crops. Nevertheless, the plants were cut when the crop height was higher than the chamber's top to allow chamber closure for measurements. Nitrous oxide ( $\text{N}_2\text{O}$ ) concentrations in the headspaces were measured for 20 minutes 4 times a day by a Megatec® IR analyzer 46i (Thermo Scientific), connected to each chamber using an automated screening system. Daily  $\text{N}_2\text{O}$  fluxes were estimated for each experimental

treatment and each day by considering the four measurements per day and the three replicates and were calculated as described previously [18]. The absence of data corresponded to periods during which the chambers were out of order or removed for technical operations on the fields (e.g., sowing, harvesting) and for undetectable N<sub>2</sub>O emissions. The data in all the remaining systems were excluded to ensure the comparison of cumulative data if measurements were missing. Cumulative N<sub>2</sub>O emissions therefore corresponded to the sum of N<sub>2</sub>O emissions observed without extrapolation, i.e., during periods for which measurements were available for all the systems simultaneously. On average, the measurements covered 60% of the length of both periods for each experiment.

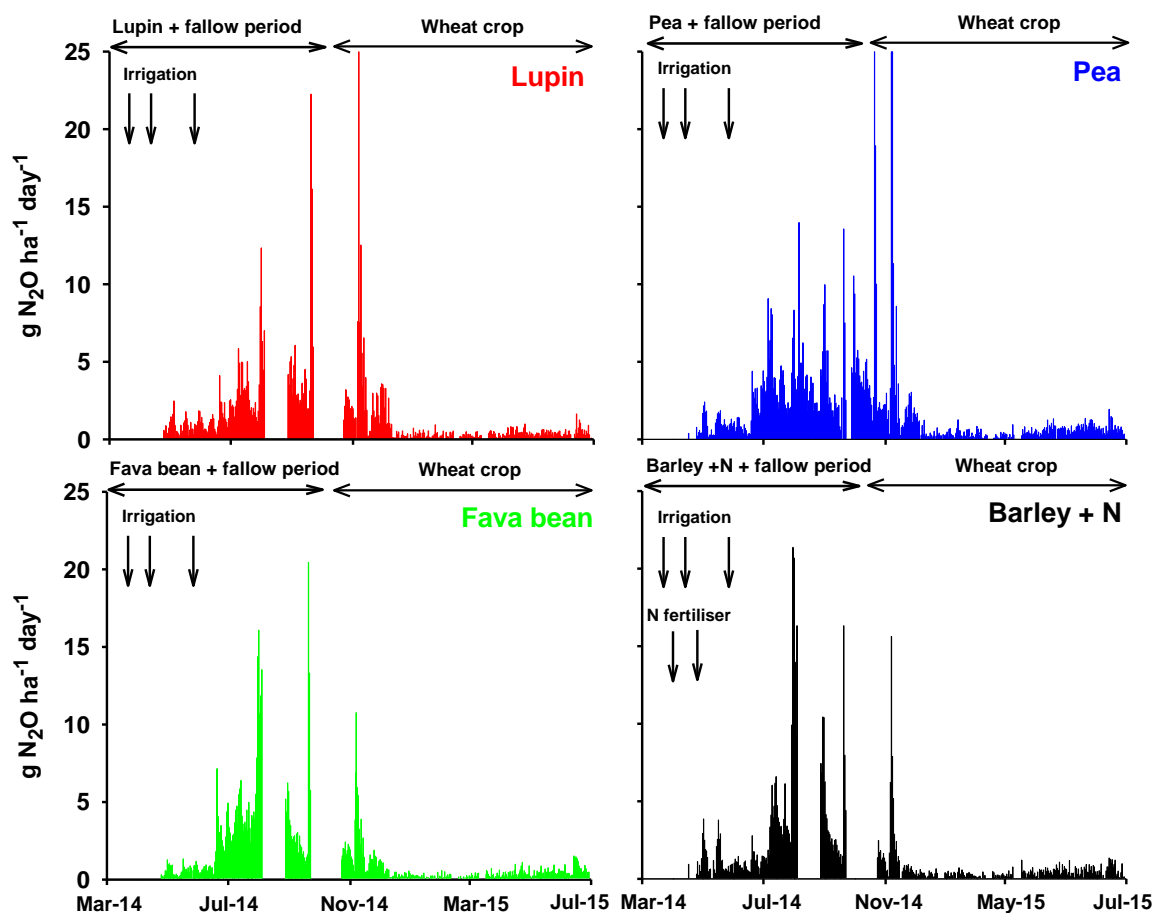
## 2.4 Statistics

Data analyses were performed using R software version 4.0.2. [19]. For each period (i.e., pre-crop + fallow and wheat crop) and each kind of crop (i.e., crops sown in March and crops sown in May), a two-way analysis of variance (ANOVA) was performed to test for a date, species and date\*species interaction effects on daily N<sub>2</sub>O emissions. Homoscedasticity and normality of variable residuals were tested with a Bartlett and Shapiro-Wilk test, respectively. A Three-factor ANOVA was performed to analyze the effects of the experiments (I and II), the kind of crops (sown in March or May), the period (pre-crop + fallow vs. wheat crop) and interactions between the three factors on cumulative values. In order to compare the effects of legumes and cereals on cumulative emissions, a one-factor ANOVA was performed separately for each kind of crop and each period. If the effects were significant, a Least significant difference" (LSD) test ( $p < 0.05$ ) was performed to compare multiple means, using the {agricolae} package.

## 3. Results

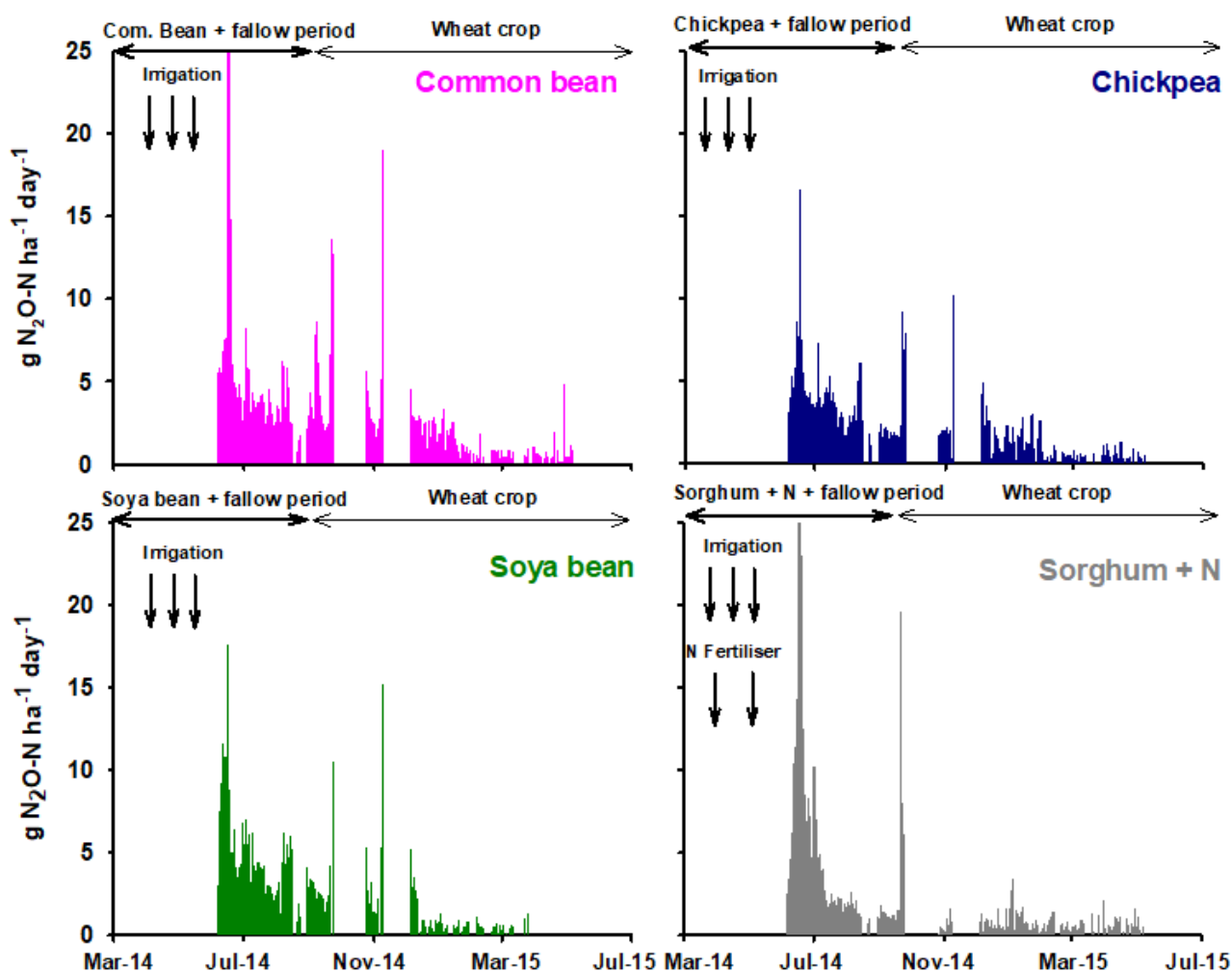
### 3.1 Daily Emissions for Experiment I

For crops sown in March 2014 and during the pre-crop + fallow period, N<sub>2</sub>O emissions varied from 0.2 to 36.0 g N ha<sup>-1</sup> d<sup>-1</sup> with an average N<sub>2</sub>O emission of 3.1 g N ha<sup>-1</sup> d<sup>-1</sup> (Figure 1). 73% of N<sub>2</sub>O emission values were lower than 2.0 g N ha<sup>-1</sup> d<sup>-1</sup>. There was no significant difference for N<sub>2</sub>O emission between pre-crops ( $P = 0.10$ ), while the effect of date was highly significant ( $P < 0.001$ ), showing that N<sub>2</sub>O emissions can present very high daily variation. Higher N<sub>2</sub>O emission peaks were measured in summer-autumn 2014, when the weather conditions were very rainy.



**Figure 1** Daily  $N_2O$  emissions for crops sown in March and for experiment I (2014-2015). Pre-crops are lupin, fava bean, pea and barley. Measurements were performed 4 times a day. Amounts of N fertilizer (for barley) and irrigated water (for all crops) are given in part 2.2.

$N_2O$  emissions during the pre-crop and fallow period for crops sown in May 2014 are presented in Figure 2. The mean  $N_2O$  emission for all treatments and periods was  $4.7 \text{ g N ha}^{-1} \text{ d}^{-1}$  with minimal and maximal  $N_2O$  emission of  $0.2$  and, exceptionally, as high as  $127.0 \text{ g N ha}^{-1} \text{ d}^{-1}$ , depending on the pre-crop and date. 50% of emission values were lower than  $2.0 \text{ g N ha}^{-1} \text{ d}^{-1}$ . The effects of pre-crop ( $P < 0.001$ ), date ( $P < 0.001$ ) and date\*pre-crop interaction ( $P < 0.001$ ) were highly significant with the highest daily emissions measured for N fertilized sorghum.



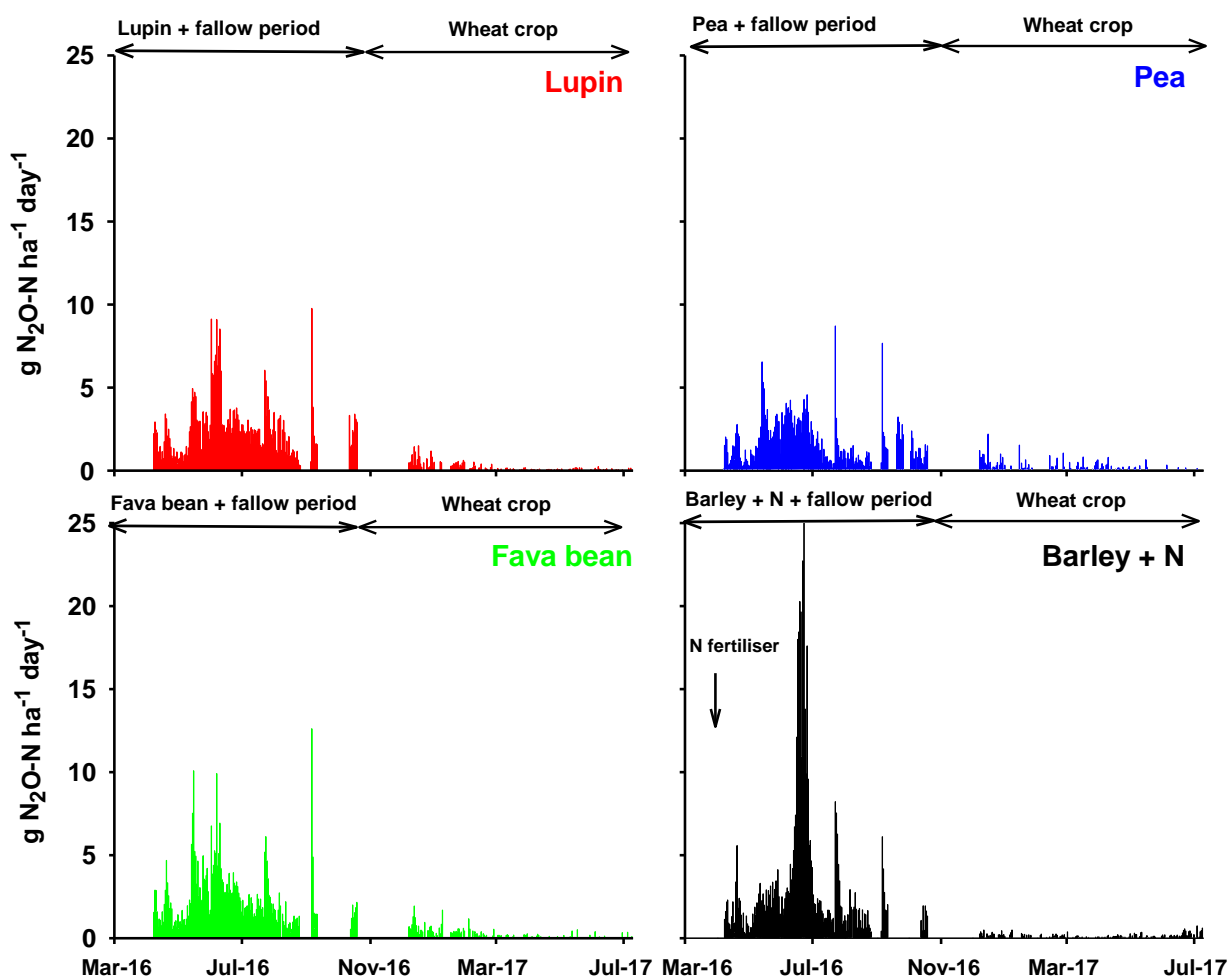
**Figure 2** Daily  $\text{N}_2\text{O}$  emissions for crops sown in May and for experiment A (2014-2015). Pre-crops are common bean, chickpea soya bean and sorghum. Measurements were performed 4 times a day. Amounts of N fertilizer (for sorghum) and irrigated water (for all crops) are given in part 2.2.

During the wheat crop period, following all pre-crops, mean  $\text{N}_2\text{O}$  emissions were respectively 1.2 and 1.4  $\text{g N ha}^{-1} \text{ d}^{-1}$  for the pre-crop sown in March and May, respectively (Figure 1 and Figure 2). Minimal and maximal emission values were respectively 0.2 and 56.9  $\text{g N ha}^{-1} \text{ d}^{-1}$  for wheat following all the pre-crops sown in March 2014, and 0.1 and 39.3  $\text{g N ha}^{-1} \text{ d}^{-1}$  for wheat following all the pre-crops sown in May 2014. The effects of date ( $P < 0.001$ ) and pre-crops ( $P < 0.001$ ) were highly significant for the pre-crops sown in March. The effects of date ( $P < 0.001$ ) and pre-crops ( $P < 0.001$ ) were equally highly significant for the pre-crops sown in May. The highest emissions were measured for wheat following legume pre-crops (mean value 1.6  $\text{g N ha}^{-1} \text{ d}^{-1}$ ) than wheat following sorghum (mean value 0.9  $\text{g N ha}^{-1} \text{ d}^{-1}$ ).

### 3.2 Daily Emissions for Experiment II

During the pre-crop and fallow period,  $\text{N}_2\text{O}$  emissions varied from 0.2 to 37.4  $\text{g N ha}^{-1} \text{ d}^{-1}$  for the crop sown in March 2016, with an average  $\text{N}_2\text{O}$  emission of 2.5  $\text{g N ha}^{-1} \text{ d}^{-1}$  (Figure 3). 67% of  $\text{N}_2\text{O}$  emissions were less than 2.0  $\text{g N ha}^{-1} \text{ d}^{-1}$  with a heterogeneous distribution of values. Higher  $\text{N}_2\text{O}$

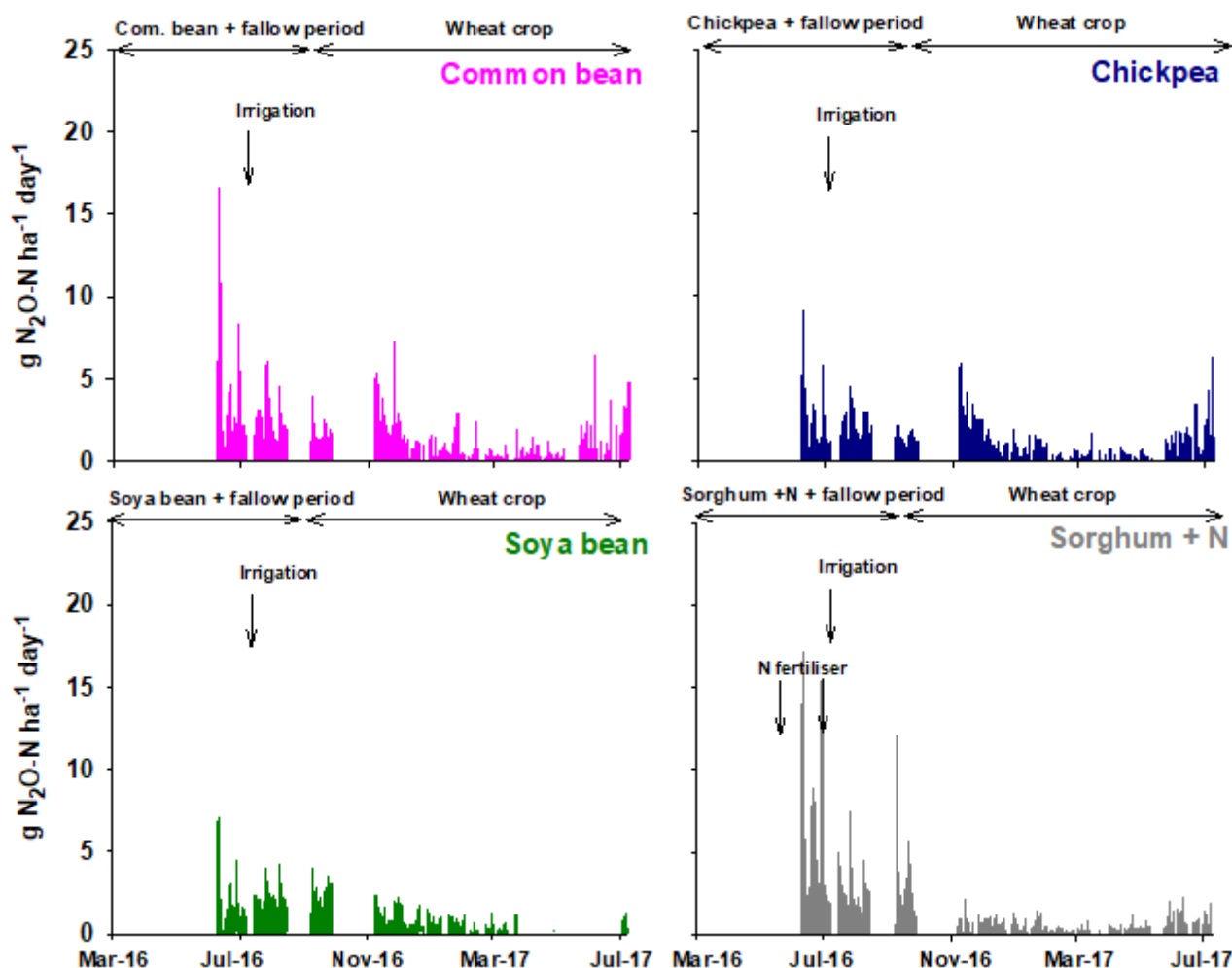
emission peaks were measured during the rainy period (spring-beginning the summer of 2016 2016). Higher peaks of N<sub>2</sub>O were measured for fertilized crops (barley or sorghum) after fertilizer inputs. The effects of date ( $P < 0.001$ ), pre-crop ( $P < 0.001$ ) and date\*pre-crop interaction ( $P < 0.001$ ) were highly significant. The main differences were observed during the period from 11 June to 30 June 2016 when higher N<sub>2</sub>O emissions were measured for barley compared to the legume pre-crops.



**Figure 3** Daily N<sub>2</sub>O emissions for crops sown in March and for experiment II (2016-2017). Pre-crops are lupin, fava bean, pea and barley. Measurements were performed 4 times a day. Amounts of N fertilizer (for barley) and irrigated water (for all crops) are given in part 2.2.

For pre-crops sown in May 2016, the mean N<sub>2</sub>O emission during the pre-crop + fallow period was 2.9 g N ha<sup>-1</sup> d<sup>-1</sup> with minimal and maximal N<sub>2</sub>O emissions of 0.1 and 32.1 g N ha<sup>-1</sup> d<sup>-1</sup>, depending on the pre-crop and date (Figure 4). 70% of emissions were less than 2.0 g N ha<sup>-1</sup> d<sup>-1</sup>. The effects of date ( $P < 0.001$ ) and pre-crop ( $P < 0.001$ ) were highly significant with higher emissions measured for sorghum compared to legume pre-crops.





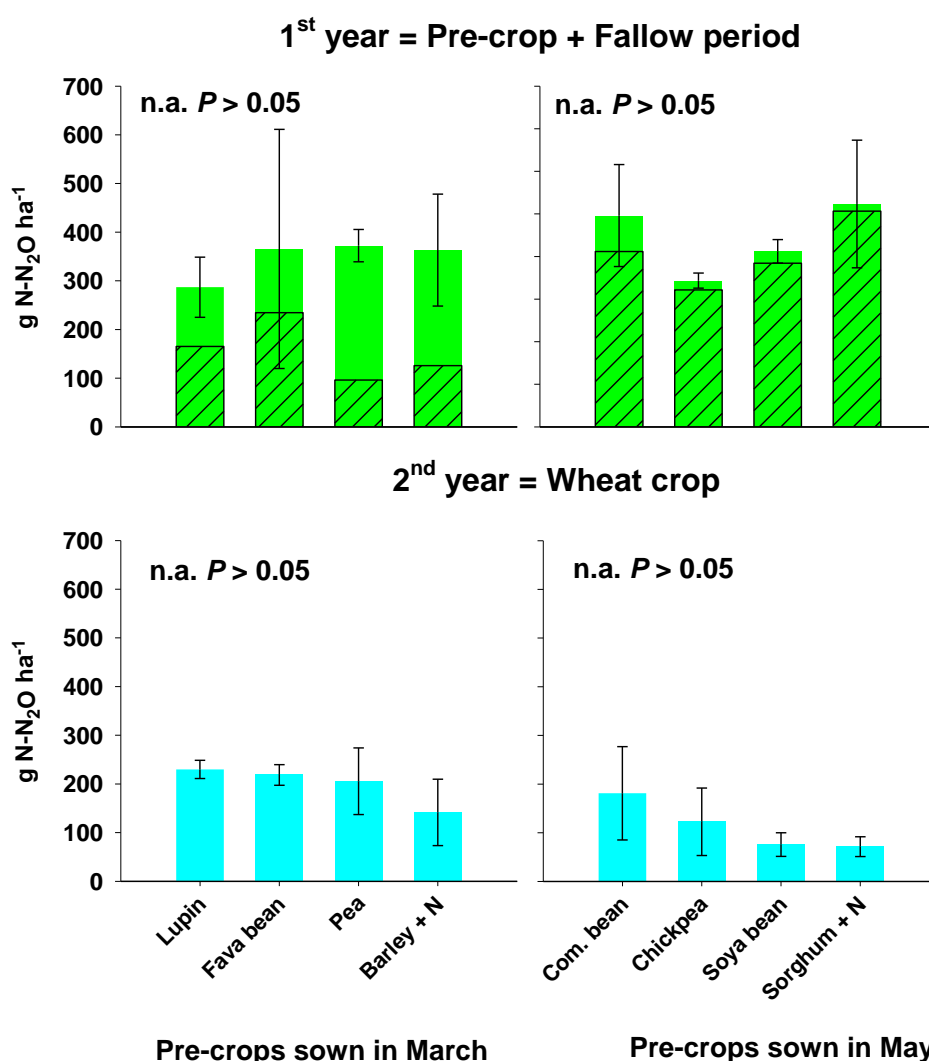
**Figure 4** Daily  $N_2O$  emissions for crops sown in May and for experiment II (2016-2017). Pre-crops are common bean, chickpea, soya bean and sorghum. Measurements were performed 4 times a day. Amounts of N fertilizer (for sorghum) and irrigated water (for all crops) are given in part 2.2.

During wheat cultivation in the second experiment, mean  $N_2O$  emissions were 0.4 and 1.2  $g N ha^{-1} d^{-1}$  for the pre-crop sown in March and May, respectively (Figure 3 and Figure 4). Minimal and maximal emission values were respectively 0.1 and 5.1  $g N ha^{-1} d^{-1}$  for all the pre-crops sown in March 2016 and 0.1 and 12.5  $g N ha^{-1} d^{-1}$  for all the pre-crops sown on May 2016. The effects of date ( $P < 0.001$ ) and pre-crops ( $P < 0.001$ ) for pre-crops sown in March and the effects of date ( $P < 0.001$ ) and pre-crops ( $P < 0.001$ ) for pre-crops sown in May were highly significant.  $N_2O$  emissions were slightly higher for wheat following common bean and chickpea compared to wheat following sorghum or Soya bean. Lower  $N_2O$  emissions were measured for wheat following barley compared to legume pre-crops (i.e., lupin, fava bean, and pea).

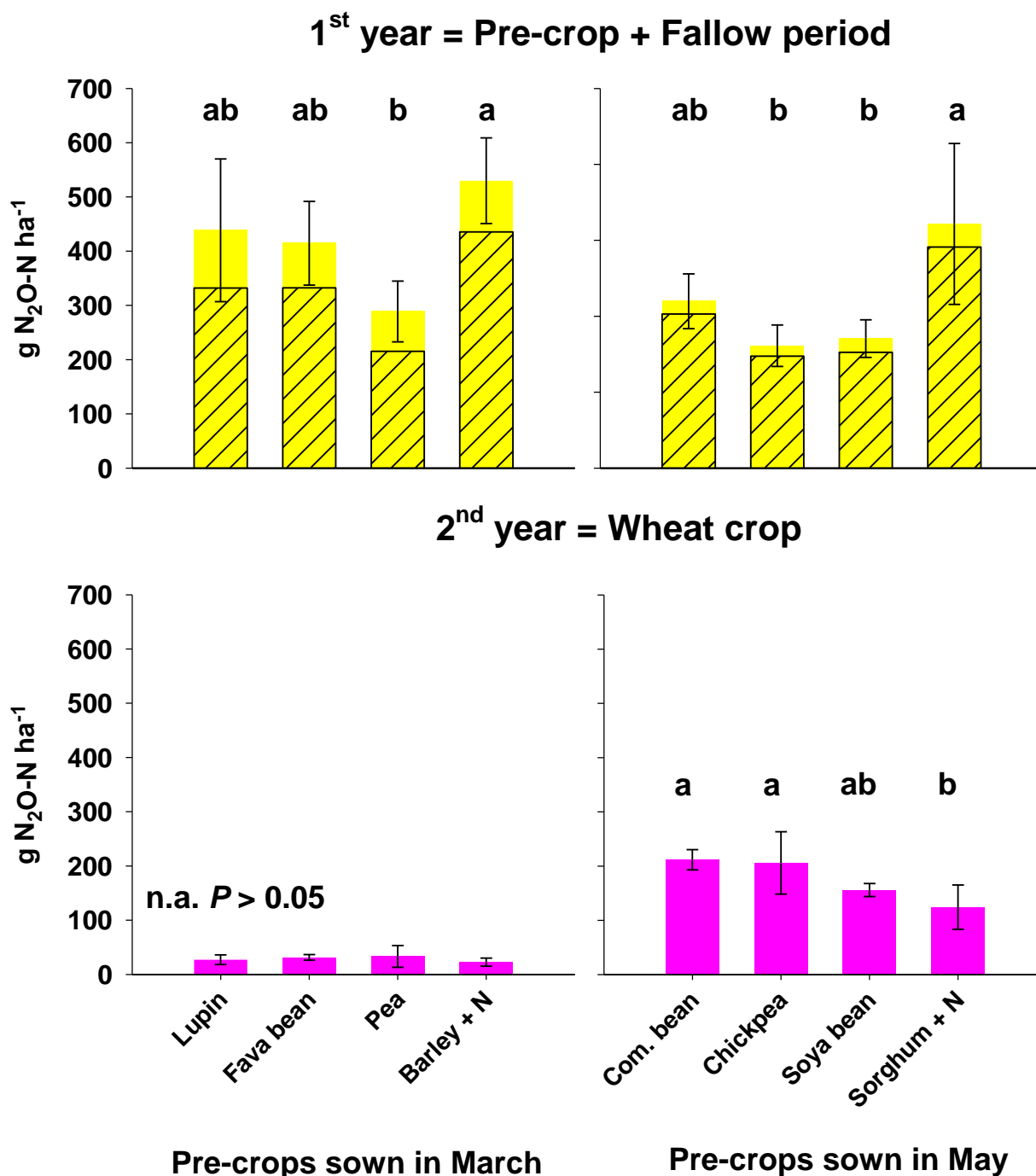
### 3.3 Cumulative Emissions for Both Experiments

Cumulative  $N_2O$  emissions for the two years of both experiments are presented in Figure 5 and Figure 6, differentiating both measurement periods. The contributions of  $N_2O$  emitted by pre-crops during the pre-crop + fallow period are also indicated. The  $N_2O$  emitted during the fallow following

the cultivation of species sown in March was higher than species sown in May, due to a long fallow period in the first case. It appears that the fallow period contributed more to N<sub>2</sub>O emitted during the pre-crop + fallow period in experiment I compared to experiment II. Emissions were significantly higher for an experiment I compared to experiment II ( $P < 0.001$ ). For both experiments, most N<sub>2</sub>O emissions occurred during the pre-crop + fallow period ( $P < 0.001$ ). The kind of crops (i.e., spring crops sown in March vs. summer crops sown in May) did not influence cumulative N<sub>2</sub>O emissions ( $P = 0.54$ ). During the pre-crop + fallow period, differences in cumulative N<sub>2</sub>O emissions between pre-crops were only observed for experiment II. Cumulative emissions for fertilized barley were significantly higher than a pea and sorghum were significantly higher than chickpea and soybean (Figure 6). During the second year (wheat crop) of experiment II significantly higher emissions were measured after legume pre-crops were sown in May (Figure 6).



**Figure 5** Cumulative emissions for crops sown in March and May for experiment I (2014-2015) and for pre-crop + fallow period, and wheat crop. Hatched bars indicate N<sub>2</sub>O emitted during legume crops. Pre-crops are lupin, fava bean, pea, barley, common bean, chickpea soya bean and sorghum. Bars for fertilized crops (barley and sorghum) are surrounded in black. LSD tests were performed separately for each period and kind of crops (i.e. crops sown in March and crops sown in May) to test the effect of pre-crops ( $P < 0.05$ ). Upper and lower bars correspond to standard deviation values.



**Figure 6** Cumulative emissions for crops sown in March and May for experiment II (2016-2017) and for both periods (pre-crop + fallow and wheat crop). Hatched bars indicate N<sub>2</sub>O emitted during legume crops. Pre-crops are lupin, fava bean, pea, barley, common bean, chickpea soya bean and sorghum. LSD tests were performed separately for each period and kind of crops (i.e. crops sown in March and crops sown in May) to test the effect of pre-crops ( $P < 0.05$ ). Upper and lower bars correspond to standard deviation values.

Considering the 2-year rotation (i.e., pre-crop + fallow period + wheat crop) the cumulative emission of Experiment I (mean value ( $\pm$  SD) =  $551.1 \pm 140.4$  g N-N<sub>2</sub>O ha<sup>-1</sup>) was significantly higher ( $P < 0.001$ ) than for Experiment II ( $429.6 \pm 100.2$  g N-N<sub>2</sub>O ha<sup>-1</sup>). The kind of crops (i.e., spring crops

sown in March vs. summer crops sown in May) did not affect cumulative N<sub>2</sub>O emissions ( $P = 0.54$ ). No significant differences ( $P > 0.10$ ) were observed between pre-crop treatments.

#### 4. Discussion

Daily N<sub>2</sub>O emissions measured during the two experiments were very low ( $<1-2.0$  g N<sub>2</sub>O ha<sup>-1</sup> d<sup>-1</sup>) in most cases. Nevertheless, daily peaks reached 130.0 g N<sub>2</sub>O ha<sup>-1</sup> d<sup>-1</sup>. N<sub>2</sub>O emissions were of the same order of magnitude as those measured at the same experimental site and in similar tilled cropping systems, with values ranging from 0.0 to 30.0-50.0 g N<sub>2</sub>O ha<sup>-1</sup> d<sup>-1</sup> [18, 20]. Although our measurements represent on average only 60% of the days in each period, N<sub>2</sub>O emissions appear to be of the same magnitude as measurements made under comparable pedoclimatic conditions in Eastern France [21] or under crop rotations including legumes [4, 22, 23]. These N<sub>2</sub>O emissions appear lower than those generally measured for different crop covers with or without legumes and with or without N (e.g. [24-26]). Likewise, cumulative N<sub>2</sub>O emissions (mean value 360.0 g N-N<sub>2</sub>O ha) during the pre-crop + fallow period, mainly due to N<sub>2</sub>O emitted during pre-crop cultivation, appear relatively low compared to other studies [4, 24, 27] and suggest that this experimental site is probably low emitting.

Indeed, soil N<sub>2</sub>O emissions vary considerably depending on many factors including pedoclimatic conditions (e.g. [24, 27, 28]), in particular soil temperature, moisture and rainfall [23, 29]. In both experiments, higher N<sub>2</sub>O emissions occurred during rainy periods. It is also generally admitted that N<sub>2</sub>O emissions originate from both anaerobic respiration such as denitrification [25, 30] and nitrification under aerobic conditions [31, 32], and could therefore occur in very variable soil moisture conditions. N fertilization (both level and form) is also a key parameter determining the amounts of N<sub>2</sub>O emitted by soils [9, 26]. For N-fertilized crops, N<sub>2</sub>O emissions are usually positively related to the quantities of nitrogen supplied [8, 28, 33]. Nevertheless, several authors [27, 34] have also shown that N<sub>2</sub>O emissions are governed together by the amount of nitrogen contained in the soil and by numerous variables (e.g., WFPS, pH, organic matter levels, tillage practices), making it difficult to compare emission levels for contrasted experimental conditions.

Numerous studies have generally observed higher N<sub>2</sub>O emissions for fertilized crops than grain legumes [4, 7, 22, 35]. Except for barley in 2016, we observed only small or no differences in N<sub>2</sub>O emissions between fertilized cereals and legumes. This was probably due (i) to the moderate or suboptimal level of N fertilization provided to cereals in environmental conditions unfavorable for high emissions and (ii) the high variability observed for cumulative emissions calculated for the pre-crop + fallow period in experiment I. Thus our first hypothesis (i.e., "N<sub>2</sub>O emissions are lower during legume crops compared to N fertilized cereals") was not systematically validated since fertilized cereals induced comparable N<sub>2</sub>O emissions compared to legume species, except for some pre-crop sown in March or May for experiment II.

Differences of N<sub>2</sub>O emissions among legume crops were slight or non-existent in this study during the pre-crop + fallow period, by studies [15] that compared N<sub>2</sub>O emissions for several grain or forage legume species. Other authors have observed differences in N<sub>2</sub>O emissions among legume species or cultivars varying according to their capacity to fix nitrogen [12]. Also, the context of low emissions for this experimental site and the relatively high variability of the measurements probably contributed to the inability to highlight differences between species.

During wheat cultivation, cumulative N<sub>2</sub>O emissions were lower than during the pre-crop period (average N<sub>2</sub>O emissions of 129.0 g N-N<sub>2</sub>O ha<sup>-1</sup>) but, surprisingly, there were no differences between treatments. The probable explanation for this lack of significant difference in the two experiments was related to (i) the context of the low-emitting experimental site and (ii) the variability of N<sub>2</sub>O emission measurements for some treatments. Our second hypothesis was therefore not validated, except for pre-crop treatments sown in May in experiment II (2016-17). After legume crops, N<sub>2</sub>O emissions are generally higher than those measured during the cultivation period [13, 36]. During the intercropping period and the following crop, pre-crop residues decompose and provide variable amounts of nitrogen to the soil through mineralization, depending on the biochemical characteristics of the residues (i.e., N content or C:N ratio) [7, 37, 38].

In this study, the amounts of N residues incorporated in the soil as well as their C:N ratio were different among legume species and cereals (Table S3), but these characteristics did not affect N<sub>2</sub>O emissions. However, in the current experiment [20], the amounts of N mineralized from crop residues were related to wheat N uptake. In addition, these authors showed by modeling that some significant amounts of nitrate were lost by leaching during the fallow period and may explain the low N<sub>2</sub>O emissions during wheat cultivation with probably insufficient amounts of nitrate remaining in the soil to induce higher N<sub>2</sub>O emissions which may in turn explain the absence of differences between experimental treatments. Under these conditions, it can be assumed that gaseous losses by denitrification (N<sub>2</sub> + N<sub>2</sub>O) are unlikely to have greatly affected the N amounts available for wheat uptake.

Based on these experimental results and taking into account the complex context of low N<sub>2</sub>O emissions from the soil, and the field experiments carried out under very contrasting climatic conditions and under-fertilized cereals, the interest in legume cultivation compared to cereals was demonstrated for only one of the field experiments concerning reducing N<sub>2</sub>O emissions by soil. This absence or weak of differences between the cumulative emissions for the different experimental treatments during the 2-year rotation is probably explained by low emission values, moderate N fertilization of cereals and variability of measurements. However, numerous studies have shown the value of introducing pulses to reduce N<sub>2</sub>O emissions across the cropping system, whether by grain legumes [7, 36], forage legumes [39, 40], cover-crop legumes [38, 40] or intercropped with a non-legume species [34, 41]. Legumes allow reducing the use of synthetic fertilizers during their growth through symbiotic fixation, and on the following crop by providing nitrogen from the mineralization of nitrogen-rich residues [42-44].

Conversely, the manufacture of synthetic fertilizers and their use in agriculture has a high impact on the use of energy resources and on the production of greenhouse gases [45]. Therefore, the introduction of legumes in cropping systems is recognized as a lever for mitigating greenhouse gas emissions [15, 46-48].

However, research is still needed to: i) clarify the origin of N<sub>2</sub>O emissions (nitrification vs. denitrification), ii) quantify these emissions under different pedoclimatic conditions, and iii) propose management methods for leguminous plants that make the best use of the nitrogen fixed during their cultivation and released when the residues are incorporated into the soil. As already implemented [23, 43, 44, 49, 50], the use of crop models is highly relevant to address these issues in future research.

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## Author Contributions

B. Nicolardot measured the N<sub>2</sub>O emissions and managed the data, M Guinet and A.S. Voisin performed the field experiments and C. Hénault managed the data. All the authors contributed to writing the paper.

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## Competing Interests

The authors have declared that no competing interests exist.

## Additional Materials

The following additional materials are uploaded at the page of this paper.

1. Figure S1: Daily rainfall (grey bars), irrigation (blue bars) and mean daily temperature (red line) for the pre-crop + fallow period and during the wheat crop in Experiment I (A) and Experiment II (B).
2. Table S1: Soil characteristics in Experiments I and II.
3. Table S2: Pre-crops and wheat characteristics and management in Experiments I and II (from [17]).
4. Table S3: Amounts, N contents and C:N ratios of pre-crops residues incorporated in soil before wheat sowing in Experiments I and II (from [17]).

## References

1. Nabuurs GJ, Mrabet R, Hatab AA, Bustamante M, Clark H, Havlík P, et al. Agriculture, forestry and other land uses (AFOLU). In: Climate change 2022: Mitigation of climate change. Contribution of working group III to the sixth assessment report of the intergovernmental panel on climate change. Cambridge and New York: Cambridge University Press; 2022. p. 185.
2. Bateman EJ, Baggs EM. Contributions of nitrification and denitrification to N<sub>2</sub>O emissions from soils at different water-filled pore space. *Biol Fertil Soils*. 2005; 41: 379-388.
3. Rochette P, Janzen HH. Towards a revised coefficient for estimating N<sub>2</sub>O emissions from legumes. *Nutr Cycl Agroecosys*. 2005; 73: 171-179.
4. Jeuffroy MH, Baranger E, Carrouée B, De Chezelles E, Gosme M, Hénault C, et al. Nitrous oxide emissions from crop rotations including wheat, oilseed rape and dry peas. *Biogeosciences*. 2013; 10: 1787-1797.

5. Zhong ZZ, Nelson LM, Lemke RL. Nitrous oxide emissions from grain legumes as affected by wetting/drying cycles and crop residues. *Biol Fertil Soils*. 2011; 47: 687-699.
6. Carrouée B, Schneider A, Flénet F, Jeuffroy MH, Nemeček T. Introduction du pois protéagineux dans des rotations à base de céréales à paille et colza: Impacts sur les performances économiques et environnementales. *Innov Agron*. 2012; 25: 125-142.
7. Uchida Y, Akiyama H. Mitigation of postharvest nitrous oxide emissions from soybean ecosystems: A review. *Soil Sci Plant Nutr*. 2013; 59: 477-487.
8. Zhong Z, Lemke RL, Nelson LM. Nitrous oxide emissions associated with nitrogen fixation by legumes. *Soil Biol Biochem*. 2009; 41: 2283-2291.
9. Li GD, Conyers MK, Schwenke GD, Hayes RC, Liu DL, Lowrie AJ, et al. Tillage does not increase nitrous oxide emissions under dryland canola (*Brassica napus* L.) in a semiarid environment of south-eastern Australia. *Soil Res*. 2016; 54: 512-522.
10. Baggs EM, Stevenson M, Pihlatie M, Rgar A, Cook H, Cadish G. Nitrous oxide emissions following application of residues and fertilizer under zero and conventional tillage. *Plant Soil*. 2003; 254: 361-370.
11. Guardia G, Abalos D, Garcia-Marco S, Quemada M, Alonso-Ayuso M, Cardenas LM, et al. Effect of cover crops on greenhouse gas emissions in an irrigated field under integrated soil fertility management. *Biogeosciences*. 2016; 13: 5245-5257.
12. Pappa VA, Rees RM, Walker RL, Baddeley JA, Watson CA. Nitrous oxide emissions and nitrate leaching in an arable rotation resulting from the presence of an intercrop. *Agric Ecosyst Environ*. 2011; 141: 153-161.
13. Wagner-Riddle C, Thurtell GW, Kidd GK, Beauchamp EG, Sweetman R. Estimated of nitrous emissions from agricultural field over 28 months. *Can J Soil Sci*. 1997; 77: 135-144.
14. Gao J, Xie Y, Jin H, Liu Y, Bai X, Ma D, et al. Nitrous oxide emission and denitrifier abundance in two agricultural soils amended with crop residues and urea in the North China plain. *PLoS One*. 2016; 11: e0154773.
15. Li GD, Schwenke GD, Hayes RC, Lowrie AJ, Lowrie RJ, Poile GJ, et al. Can legume species, crop residue management or no-till mitigate nitrous oxide emissions from a legume-wheat crop rotation in a semi-arid environment? *Soil Tillage Res*. 2021; 209: 104910.
16. Food and agriculture organization of the United Nations. World reference base for soil resources 2014. International soil classification system for naming soils and creating legends for soil maps. Rome: FAO; 2014. p. 192.
17. Guinet M, Nicolardot B, Voisin AS. Nitrogen benefits of ten legume pre-crops for wheat assessed by field measurements and modelling. *Eur J Agron*. 2020; 120: 126151.
18. Vermue A, Phillipot L, Munier-Jolain N, Hénault C, Nicolardot B. Influence of integrated weed management system on N-cycling microbial communities and N<sub>2</sub>O emissions. *Plant Soil*. 2013; 373: 501-514.
19. R Core Team. A language and environment for statistical computing [Internet]. Vienna, Austria: R foundation for statistical computing; 2020. Available from: <https://www.r-project.org/>.
20. Vermue A, Nicolardot B, Hénault C. High N<sub>2</sub>O variations induced by agricultural practices in integrated weed management systems. *Agron Sustain Dev*. 2016; 36: 45.
21. Hénault C, Devis X, Page S, Justes E, Reau R, Germon JC. Nitrous oxide emissions under different soil and land management conditions. *Biol Fertil Soils*. 1998; 26: 199-207.

22. Ma Y, Schwenke G, Sun LY, Liu DL, Wang B, Yang B. Modeling the impact of crop rotation with legume on nitrous oxide emissions from rain-fed agricultural systems in Australia under alternative future climate scenarios. *Sci Total Environ*. 2018; 630: 1544-1552.
23. Barton L, Butterbach-Bahl K, Liese R, Murphy DV. Nitrous oxide fluxes from a grain-legume crop (narrow-leaved lupin) grown in a semiarid climate. *Glob Chang Biol*. 2011; 17: 1153-1166.
24. Eichner ML. Nitrous oxide emissions from fertilized soils: Summary of available data. *J Environ Qual*. 1990; 19: 272-280.
25. Dobbie KE, Smith KE. Nitrous oxide emission factors for agricultural soils in Great Britain: The impact of soil water-filled pore space and other controlling variables. *Glob Change Biol*. 2003; 9: 204-218.
26. Officer SJ, Phillips F, Kearney G, Armstrong R, Graham J, Partington D. Response of soil nitrous oxide flux to nitrogen fertiliser application and legume rotation in a semi-arid climate, identified by smoothing spline models. *Soil Res*. 2015; 53: 227-241.
27. Jensen ES, Peoples MB, Boddey RM, Gresshoff PM, Hauggaard-Nielsen H. Legumes for mitigation of climate change and the provision of feedstock for biofuels and biorefineries. A review. *Agron Sustain Dev*. 2012; 32: 329-364.
28. Le Gall C, Jeuffroy MH, Hénault C, Python Y, Cohan JP, Parnaudeau V, et al. Analyser et estimer les émissions de N<sub>2</sub>O dans des systèmes de culture français. *Innov Agron*. 2014; 12: 97-112.
29. Schaufler G, Kitzler B, Schindlbacher A, Skiba U, Sutton A, Zechmeister-Boltenstern A. Greenhouse gas emissions from European soils under different land use: Effect of soil moisture and temperature. *Eur J Soil Sci*. 2010; 61: 683-696.
30. Linn DM, Doran JW. Effect of water-filled pore space on carbon dioxide and nitrous oxide production in tilled and non-tilled soils. *Soil Sci Soc Am J*. 1984; 48: 1267-1272.
31. Mathieu O, Hénault C, Lévêque J, Baujard E, Milloux MJ, Andreux F. Quantifying the contribution of nitrification and denitrification to the nitrous oxide flux using <sup>15</sup>N tracers. *Environ Pollut*. 2006; 144: 933-940.
32. Hénault C, Grossel A, Mary B, Roussel M, Léonard J. Nitrous oxide emission by agricultural soils: A review of spatial and temporal variability for mitigation. *Pedosphere*. 2012; 22: 426-433.
33. Schwenke GD, Brock PM, Haigh BM, Herridge DF. Greenhouse gas emission reductions in subtropical cereal-based cropping sequences using legumes, DMPP-coated urea and split timings of urea application. *Soil Res*. 2018; 56: 724-736.
34. Senbayram M, Wenthe C, Lingner A, Isselstein J, Steinmann H, Kaya C, et al. Legume-based mixed intercropping systems may lower agricultural born N<sub>2</sub>O emissions. *Energy Sustain Soc*. 2015; 6: 2.
35. Reckling M, Bergkvist G, Watson CA, Stoddard FL, Zander PM, Walker RL, et al. Trade-offs between economic and environmental impacts of introducing legumes into cropping systems. *Front Plant Sci*. 2016; 7: 669.
36. Véricel G, Dubois S, Duval R, Flénet F, Fontaine L, Fourrié L, et al. Impact de l'introduction des légumineuses dans les systèmes de culture sur les émissions de protoxyde d'azote. *Innov Agron*. 2018; 63: 211-229.
37. Huang Y, Zou J, Zheng X, Wang X, Wang Y, Xu X. Nitrous oxide emissions as influenced by amendment of plant residues with different C:N ratios. *Soil Biol Biochem*. 2004; 36: 973-981.



38. Gomes J, Bayer C, de Souza Costa F, de Cássia Piccolo M, Zanatta JA, Vieira FCB, et al. Soil nitrous oxide emissions in long-term cover crops-based rotations under subtropical climate. *Soil Tillage Res.* 2009; 106: 36-44.
39. Van der Weerden TJ, Sherlock RR, Williams PH, Cameron KC. Nitrous oxide and methane oxidation by soil following cultivation of two different leguminous pastures. *Biol Fertil Soils.* 1999; 30: 52-60.
40. Basche AD, Miguez FE, Kaspar TC, Castellano MJ. Do cover crops increase or decrease nitrous oxide emissions? A meta-analysis. *J Soil Water Conserv.* 2014; 69: 471-482.
41. Westphal M, Tenuta M, Entz MH. Nitrous oxide emissions with organic crop production depends on fall soil moisture. *Agric Ecosyst Environ.* 2018; 254: 41-49.
42. Crews TE, Peoples MB. Legume versus fertilizer sources of nitrogen: Ecological tradeoffs and human needs. *Agric Ecosyst Environ.* 2004; 102: 279-297.
43. De Antoni Migliorati M, Parton WJ, Del Grosso SJ, Grace PR, Bell MJ, Strazzabosco A, et al. Legumes or nitrification inhibitors to reduce N<sub>2</sub>O emissions from subtropical cereal cropping systems in oxisols. *Agric Ecosyst Environ.* 2015; 213: 228-240.
44. Mielenz H, Thorburn PJ, Harris RH, Officer SJ, LI GD, Schwenke GD, et al. Nitrous oxide emissions from grain production systems across a wide range of environmental conditions in eastern Australia. *Soil Res.* 2016; 54: 659-674.
45. Peoples MB, Hauggaard-Nielsen H, Jensen ES. The potential environmental benefits and risks derived from legumes in rotations. In: *Nitrogen fixation in crop production*. New York: John Wiley & Sons, Inc.; 2009. pp. 349-385.
46. Mahama GY, Prasad PVV, Roozeboom KL, Nippert JB, Rice CW. Reduction of nitrogen fertilizer requirements and nitrous oxide emissions using legume cover crops in a no-tillage sorghum production system. *Sustainability.* 2020; 12: 4403.
47. Oliveira M, Castro C, Coutinho J, Trindade H. Grain legume-based cropping systems can mitigate greenhouse gas emissions from cereal under Mediterranean conditions. *Agric Ecosyst Environ.* 2021; 313: 107406.
48. Saha D, Kaye JP, Bhowmik A, Bruns MA, Wallace JM, Kemanian AR. Organic fertility inputs synergistically increase denitrification-derived nitrous oxide emissions in agroecosystems. *Ecol Appl.* 2021; 31: e2403.
49. Plaza-Bonilla D, Nogue-Serra I, Raffailac D, Cantero-Martinez C. Carbon footprint of cropping systems with grain legumes and cover crops: A case-study in SW France. *Agric Syst.* 2018; 167: 92-102.
50. Plaza-Bonilla D, Léonard J, Peyrard C, Mary B, Justes E. Precipitation gradient and crop management affect N<sub>2</sub>O emissions: Simulation of mitigation strategies in rainfed Mediterranean conditions. *Agric Ecosyst Environ.* 2017; 238: 89-103.